



Monoterpene emission of *Quercus suber* L. highly infested by *Cerambyx welensii* Küster

Israel Sánchez-Osorio¹ · Gloria López-Pantoja¹ · Raúl Tapias¹ · Evangelina Pareja-Sánchez² · Luis Domínguez¹

Received: 25 February 2019 / Accepted: 27 August 2019 / Published online: 18 October 2019
© INRA and Springer-Verlag France SAS, part of Springer Nature 2019

Abstract

• **Key message** Cork oaks highly infested by *Cerambyx welensii* emit an amount of limonene at dusk, when *C. welensii* adults become active. In contrast, emissions by neighboring cork oaks free of *C. welensii* are dominated by pinene-type compounds.

• **Context** The activity of the woodborer *Cerambyx welensii* Küster is a key factor in the decline of *Quercus suber* L. dehesas.

• **Aims** This study aimed to estimate whether trees highly infested by *C. welensii* exhibited a peculiar emission profile, with known antennally active compounds.

• **Methods** Monoterpenes were sampled in situ in 2006 (day/late evening) and 2008 (early evening) from *Q. suber* stratified by whether or not trees were highly infested by *C. welensii* and analyzed by gas chromatography.

• **Results** Limonene, α -pinene, β -pinene, sabinene, and myrcene accounted for over 87.2% of overall monoterpene emissions. Monoterpene composition and emission rates differed between the two groups, both during daytime and early evening, with a high presence of limonene in infested trees and dominance of pinene-type compounds in non-infested trees.

• **Conclusion** This work evidenced differences in foliar monoterpene emissions between *Q. suber* trees highly infested by *C. welensii* and non-infested trees, with a high presence of limonene in the former and dominance of pinene-type compounds in non-infested trees. We hypothesize that the detection—especially during the onset of insects daily flight—of certain compounds (e.g., limonene), together with the detection of specific ratios of several monoterpenes (e.g., those of limonene to pinene-type compounds), has a role in the intraspecific host selection by *C. welensii*.

Keywords Dehesa · Woodborer · Limonene · Pinene-type · Quercus decline

Handling Editor: Aurélien Sallé

Contribution of the co-authors Project design, I. Sánchez-Osorio. Data collection, all authors. Data analysis and paper writing, I. Sánchez-Osorio. All authors read and approved the manuscript.

This article is part of the topical collection on *Entomological issues during forest diebacks*

✉ Israel Sánchez-Osorio
isanchez@uhu.es

¹ Departamento de Ciencias Agroforestales, ETSI La Rábida, University of Huelva, 21819 Palos de la Frontera, Huelva, Spain

² Departament de Producció Vegetal i Ciència Forestal (EEAD-CSIC Associated Unit), University of Lleida, Lleida, Spain

1 Introduction

Open Mediterranean woodlands (called *dehesas* in Spain) are outstanding agroforestry systems, protected under the European Habitats Directive (Council Directive 92/43/EEC). There are 4 million ha of this type of woodlands in the western Iberian Peninsula, with *Quercus suber* L. the dominant tree species covering 716,000 ha in Portugal and 300,000 ha in Spain; together, these areas produce about two-thirds of the world cork production (Alejano et al. 2011; ICNF 2013). Cork oak stands are threatened by multiple stresses, both anthropogenic and natural (Aronson et al. 2009), being the activity of wood-boring insects, such as *Cerambyx welensii* Küster, the key factor in the decline of *Q. suber* in south-west Spain (Sallé et al. 2014; Tiberi et al. 2016; Torres-Vila et al. 2017). *C. welensii* is a

large (up to 60 mm long) cerambycid that affects several deciduous trees, mainly within the *Quercus* genus (Vives 2000). The flight period of *C. welensii* spans from mid-May to late-July. Adults become active at dusk, with peak adult activity occurring mainly between 21:00 and 24:00 h (Lpez-Pantoja et al. 2008; data of the authors), though there are also reports of adult activity during the daytime (Vives 2000; Torres-Vila et al. 2016). Adverse impact of *C. welensii* adults on tree health has not been reported; however, larvae bore into wood causing tree branches and trunks to break (Lpez-Pantoja et al. 2008) and facilitate infection by plant pathogens and wood-decaying fungi (Martín et al. 2005).

Terpenes and other volatile organic compounds play important roles in both plant physiology and interactions of plants with their environment, e.g., protecting plants against stressors or acting as a mechanism for interplant communication (Loreto et al. 2014). There is considerable evidence that longhorn beetles are attracted to volatiles, such as monoterpenes, emitted by plants. In some species, host volatiles synergize response to sex pheromones (Allison et al. 2004; Millar and Hanks 2017). Little is known, however, about the role of such compounds in host location by wood-boring beetles that infest non-coniferous trees (see, for example, Millar and Hanks 2017). Among oaks, *Q. suber* has been widely reported to be a strong monoterpene emitter (Staudt et al. 2004; Pio et al. 2005; Lavoit et al. 2011), its emissions showing both seasonal and site variations that can be attributed to microclimate conditions (Staudt et al. 2004) as well as genetic diversity (Staudt et al. 2004; Loreto et al. 2014). In addition, biotic and environmental stresses may affect monoterpene emission, since monoterpene precursors are derived from photosynthetic activity (Núñez et al. 2002; Lavoit et al. 2011). Despite this variability, the emission profile of a given *Q. suber* tree is considered to be relatively stable (Staudt et al. 2004), consisting of four dominant monoterpenes: α -pinene, β -pinene, sabinene, and limonene, and a minor presence of myrcene (Staudt et al. 2004; Pio et al. 2005; Lavoit et al. 2011).

Observations indicate that *C. welensii* prefers to colonize weak, damaged, or old trees and that visual cues play a role in host location (Lpez-Pantoja et al. 2008; Torres-Vila et al. 2017). In addition, the cork harvesting (every 9 years in south-west Spain) frequently causes bark damage to trees, facilitating oviposition of *C. welensii* as well as triggering the release of volatiles that may be attractive to *C. welensii* (Snchez-Osorio et al. 2016). Moreover, interannual presence of large numbers of *C. welensii* on the same trees and aggregation phenomena in trees showing bark exudates have been reported (Lpez-Pantoja et al. 2008). In field experiments, *C. welensii* has been found to be attracted to traps baited with a mixture of red wine, vinegar, and sugar (Torres-Vila et al. 2012) or with synthetic volatiles mimicking fermenting plant material (Snchez-Osorio et al. 2016). *C. welensii* antennae

respond to limonene, myrcene, and pinene-type compounds (Sánchez-Osorio 2005), and field catches obtained with synthetic volatiles at a low release rate (1.2 g day^{-1}) slightly improved when β -pinene was added to the bait (Snchez-Osorio et al. 2016).

The aim of this study was to find out whether the foliar emissions of monoterpenes by *Q. suber* mediate intraspecific host selection by *C. welensii*. For this purpose, we analyzed monoterpene emission during daytime, early evening, and late evening by *Q. suber* trees stratified by whether they were or not highly infested by *C. welensii*. Time spans studied were chosen according to the daily flight activity of *C. welensii*, including the early evening. We hypothesized that (i) the monoterpene emission of *Q. suber* would exhibit small-scale spatial and temporal variations and (ii) there would be a relationship between emission patterns and infestation by *C. welensii*. Identification of the factors related to host recognition, in particular, volatile compounds released by trees, might facilitate the development of lures and ultimately sustainable pest management strategies, for example, the use of baited traps for detection, monitoring, and, eventually, mass trapping of adults.

2 Materials and methods

2.1 Plant material

Foliar monoterpene samples were collected in 2006 and 2008 in a dehesa located in south-west Spain ($37^{\circ} 15' 43.73'' \text{ N}$, $6^{\circ} 28' 34.65'' \text{ W}$, 80 m a.s.l.) strongly affected by *C. welensii* and composed primarily of *Q. suber* trees (density of 75 trees ha^{-1} ; mean perimeter at breast height was 168.4 and 131.1 cm, infested and non-infested trees, respectively). The mean temperature for June was similar in both years ($23^{\circ} \text{ C} \pm 0.6$), while the annual precipitation was 819 mm in 2006 (with 30 mm of rainfall in June) and 588 mm in 2008 (with no measurable rainfall in June) (Almonte meteorological station $37^{\circ} 08' 53'' \text{ N}$, $6^{\circ} 28' 35'' \text{ W}$).

We studied 12 trees in 2006, and the same 12 plus an additional 24 (total of 36 trees) in 2008. Based on both sightings of *C. welensii* adults on each tree between 2002 and 2007 (Lpez-Pantoja et al. 2008) and the presence of symptoms that could indicate colonization by *C. welensii* larvae (exit holes and sawdust), two groups of trees were established each year. The first group ($n = 9$ in 2006; $n = 18$ in 2008), hereafter referred to as infested trees, was composed of the trees with the higher occurrence of adults (mean of 54 ± 23 insects per tree between 2002 and 2007). These trees also showed the greatest amount of damage by large woodborers and systematically showed symptoms that could indicate colonization by *C. welensii* larvae, though the presence of larvae was not verified. The second group ($n = 3$ in 2006; $n = 18$ in

2008), hereafter referred to as non-infested trees, was composed of neighboring *Q. suber* on which neither adults, nor damage, nor symptoms attributable to this woodborer were observed (and this also stands after a survey carried out in 2008).

Monoterpene volatiles were collected in the field from intact twigs (two twigs per tree) in 2006 and from pruned twigs (one twig per tree, sampled immediately after pruning) in 2008. All twigs sampled (30 cm long, 53–187 leaves; mean dry weight of 6.2 ± 2.3 g in 2006 and 14.48 ± 0.7 g in 2008) were selected from near the tip of sun-exposed branches, at a height of ≈ 2 m, and in the 2006 samplings, the pairs of twigs from the same tree were 1 m apart.

2.2 Monoterpene sampling and analysis

Based on flight period estimates by Lopez-Pantoja et al. (2008) for *C. welensii* in the studied dehesa habitat, monoterpenes were sampled near the peak of the *C. welensii* flight period each year. In 2006, volatiles were sampled during two different periods: the daytime period (13:00 to 14:30 h, June 12) close to the daily peak of *Q. suber* emissions (Pio et al. 2005), and the late evening period (21:30 to 23:00 h, June 13), from the same twigs as those studied on 12 June, during the peak activity of *C. welensii* adults (Lopez-Pantoja et al. 2008). To study short-term variations in monoterpene emissions during a prior period that included the onset of *C. welensii* daily flight, in 2008, samples were collected five times over the early evening period (19:00, 19:35, 20:10, 20:45, and 21:20 h, June 24). For this purpose, three to four different trees from each study group were randomly selected for analysis in each measurement time (Table 1).

Monoterpenes were sampled using a custom-made head-space collection system with bag enclosure (Maja et al. 2014), based on the “aeration system” described by Zhang et al. (1999). Each collection system consisted of two Teflon sampling lines connected to a diaphragm pump (SP 200 EC-LC; Schwarzer Precision, Essen, Germany); the two lines allowed for simultaneous collection of a sample from a twig and a blank control sample (from a measurement chamber in which no twig had been enclosed). The measurement chambers

consisted of 38 cm \times 25 cm polyester oven bags (Albal, Cofresco, Madrid, Spain). Each twig sampled was enclosed in a new measurement chamber, with the bag opening fastened around the stem with garden wire; a Teflon air-inlet attached to a charcoal filter glass tube (Split Vent Trap containing 1.2 g charcoal; Cromlab, Barcelona, Spain) allowed air to enter the chamber. Monoterpenes were trapped by drawing the air inside each bag through a glass sorbent tube (150 mg 403 Orbo Tenax TA tube, 60/80 mesh; Sigma-Aldrich, Madrid, Spain) for 5 min at a flow rate of 120 ml min⁻¹. All sample tubes were sealed with Teflon caps immediately after collection, kept at ≈ 4 °C and taken to the laboratory where samples were stored at -28 °C until analysis (in the following 24–48 h). Temperature and relative humidity, besides photosynthetic photon flux density (PPFD) in 2008, were monitored during samplings.

Sample tubes were spiked with 1-bromo-2-chlorobenzene as an internal standard and eluted with 1.5 ml cyclohexane (purity > 99.5% and > 99%, respectively; Sigma-Aldrich, Madrid, Spain). Monoterpene analyses were performed with 1- μ l aliquots of the cyclohexane solution either by gas chromatography-mass spectrometry (GC-MS) (GC type 6890N, MSD 5973; Agilent, Santa Clara, USA) in 2006 or by gas chromatography with a flame ionization detector (Agilent 6890N GC system) in 2008. In both cases, an HP-5MS column (0.25 mm \times 30 m \times 0.25 μ m) using helium as the carrier gas (1 ml min⁻¹) was employed, and the oven temperature was programmed with the following conditions: the initial temperature (46 °C) was increased at 30 °C min⁻¹ to 70 °C, held steady for 4 min, and then increased at rates of 5 °C min⁻¹ to 80 °C, 4.5 °C min⁻¹ to 90 °C, and 50 °C min⁻¹ to 300 °C.

Peaks were identified by comparison with pure standards (Sigma-Aldrich; Madrid, Spain; purity $\geq 94\%$) and with mass spectra in the NIST 02 library (MSD Chemstation Build 75 software). For quantitative analysis, calibration curves were determined for α -pinene, β -pinene, sabinene, limonene, and myrcene (0.05–1 ppm solutions). The calibration curves were highly significant ($R^2 > 0.99$) in all cases. Emission rates were calculated in nanograms of compound per gram of dry weight of leaf material per hour (ng gdw⁻¹ h⁻¹), after subtracting the values measured in blank chambers. We analyzed only the monoterpenes, other volatile compounds (leaf alcohols, sesquiterpenes, etc.) that may have been emitted by the foliage not being considered here.

2.3 Statistical analyses

Comparative analyses of monoterpene emissions were performed using linear mixed models (LMMs, lme4 package; Bates et al. 2014), for the total emission variable (T_{em} , the sum of the five main compounds: α -pinene, β -pinene, sabinene, limonene, and myrcene), and for each of the main

Table 1 Factorial treatment combinations and number of replicates, to study short-term monoterpene emissions of *Quercus suber* trees over the early evening period

Infestation	Measurement time (h)					Total
	19:00	19:35	20:10	20:45	21:20	
Infested	4	4	3	4	3	18
Non-infested	3	4	4	3	4	18
Total	7	8	7	7	7	36

monoterpenes separately. Mean monoterpene emission was treated as the dependent variable (the emission from each tree in 2006 being the mean of values obtained from the two twigs studied). Infestation by *C. welensii* (used as a class variable) and measurement time (for the 2008 dataset) were used as fixed factors (Table 2). Tree identity (intercept model) was included as a random factor. The significance of the main factors was assessed using likelihood ratios. For multiple testing, Benjamini-Hochberg adjusted *P* values were used (lsmmeans package; Lenth 2014). Within-tree repeatability of emissions (intraclass correlation, *r*) (Falconer and Mackay 1996) was analyzed both between twigs from each tree and over time.

We performed permutational multivariate analysis of variance (PERMANOVA, using the Bray-Curtis dissimilarity index and 999 permutations) and multivariate dispersion analysis (vegan package; Oksanen et al. 2015) to determine whether there were differences in emission composition between trees infested by *C. welensii* and trees not infested by this woodborer. The variables included in the PERMANOVA were individual relative emissions (percentage of T_{em}) for limonene and myrcene, as well as a newly created variable (called pinene-type) corresponding to the sum of relative emissions for α -pinene, β -pinene, and sabinene. This pinene-type variable could aid us to differentiate a pinene-type emission profile, as α -pinene, β -pinene, and sabinene are highly correlated (positive) in *Q. suber* (Staudt et al. 2004). Log-likelihood ratio tests (G-tests) (with Williams' correction) were used to determine whether the percentage of trees with a limonene profile differed between infested and non-infested *Q. suber*.

All statistical analyses were performed in R software, version 3.1.0, using $\alpha = 0.05$ and $\alpha = 0.1$ as the thresholds for significance and marginal significance, respectively.

3 Results

3.1 Monoterpene emission by *Quercus suber* leaves

The air temperature (*T*) was high across the three sampling times, especially throughout 2006 daytime and 2008 early evening samplings (> 28.1 °C). Relative humidity (RH) was very low across those two sampling periods (< 40% overall). Further, the photosynthetic photon flux density (PPFD) sharply decreased over the 2008 early evening sampling (from 1097 to 75 $\mu\text{mol m}^{-2} \text{s}^{-1}$) and reached values under 45 $\mu\text{mol m}^{-2} \text{s}^{-1}$ in the 2006 late evening sampling (Fig. 1).

The main monoterpenes emitted by *Q. suber* in 2006 and 2008 were limonene, α -pinene, β -pinene, sabinene, and myrcene (Table 1; Fig. 2); these five compounds accounted for 87.2 to 95.6% of the overall monoterpene emission by each tree (2006 daytime and 2008 early evening, respectively) but accounted for 100% of all monoterpenes detected from five infested trees and one non-infested tree in the 2006 late evening sampling. During the 2006 late evening sampling, two infested trees and one non-infested tree showed no measurable emissions. Low emissions ($\leq 2\%$ of the overall emission by each tree in 2006 daytime, and traces in both 2006 late evening and 2008 early evening samples) were found for cineole, camphene, γ -terpinene, α -terpinene, ρ -cymene, α -phellandrene, and α -thujene. Given these results, only the emissions of limonene, α -pinene, β -pinene, sabinene, and myrcene were considered

Table 2 Leaf monoterpene emissions (mean \pm SE), expressed as total monoterpene emission rate (T_{em} , the emission of the five main monoterpenes pooled), individual monoterpene relative emission, and

abundance of trees with a limonene-type profile, of *Quercus suber* trees in 2006 and 2008 in south-west Spain. Trees were stratified by whether or not they were infested by *Cerambyx welensii*

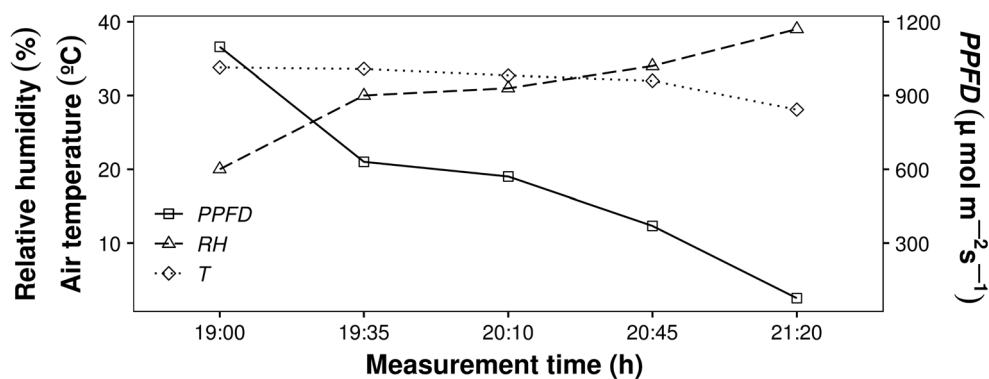
Samp [#] infestation	Total emission rate (T_{em} , ng gdw ⁻¹ h ⁻¹)	Emission composition (sum = 100%)					Limonene profile* (% trees)
		α -Pinene	β -Pinene	Sabinene	Myrcene	Limonene	
Daytime (12 June 2006; 13:00–14:30 h)							
Infested	1950.22 \pm 339.06a	12.5 \pm 1.6a	9.3 \pm 1.2a	22.7 \pm 2.7a	7.1 \pm 0.4a	48.4 \pm 5.0a	88.9
Non-infested	1510.21 \pm 541.74a	22.4 \pm 2.5b	17.1 \pm 1.2b	43.9 \pm 2.8b	5.3 \pm 0.5a	11.3 \pm 5.9b	0
Late evening (13 June 2006; 21:30–23:00 h)							
Infested	26.73 \pm 2.92a	22.4 \pm 13.9a	2.5 \pm 2.5a	1.8 \pm 1.8a	0	73.3 \pm 15.1a	85.7
Non-infested	25.22 \pm 2.61a	22.7 \pm 19.1a	0 a	0 a	0	77.3 \pm 19.1a	66.6
Early evening (24 June 2008; 19:00–21:20 h)							
Infested	962.01 \pm 197.06a	11.2 \pm 3.1a	6.4 \pm 2.2a	15.9 \pm 4.1a	0.7 \pm 0.4a	65.8 \pm 8.5a	88.9
Non-infested	1209.52 \pm 562.06a	21.2 \pm 5.1a	6.3 \pm 1.9a	19.6 \pm 5.1a	0.5 \pm 0.2a	52.5 \pm 10.4a	58.8

Different letters indicate significant differences by infestation status within each sampling time (LMM followed by likelihood ratio tests, $\alpha = 0.05$)

[#] Sampling time

*Limonene percentage > 30% of T_{em} (Staudt et al. 2004)

Fig. 1 Air temperature (T), relative humidity (RH), and photosynthetic photon flux density ($PPFD$) during sampling of monoterpene emissions from *Quercus suber* in 2008



for further statistical analysis, and thereafter, the T_{em} variable refers to the sum of observed emissions of these five compounds. Within-tree repeatability of monoterpene relative emissions (intra-class correlation, r) was high both between twigs from each tree in daytime samples (r ranging from 0.91 to 0.98, α -pinene and limonene, respectively) and between 2006 daytime and 2008 early evening samples (r ranging from 0.62 to 0.94, α -pinene and limonene, respectively). In contrast, repeatability of emission rates was low ($r < 0.3$ overall) between twigs from each tree, as did over both short-time (in 2006) and long-time (between 2006 daytime and 2008 early evening samples) periods.

The T_{em} variable reached a mean of $1840.2 \text{ ng gdw}^{-1} \text{ h}^{-1}$ in the 2006 daytime sampling and ranged between 190 and $2450 \text{ ng gdw}^{-1} \text{ h}^{-1}$ (21:20 and 20:10 h, respectively) in the 2008 early evening sampling; furthermore, T_{em} was very low in the 2006 late evening sampling (average of $26 \text{ ng gdw}^{-1} \text{ h}^{-1}$). The overall emission composition of trees showed a greater presence of limonene in both early and late evening than daytime emissions (Table 1). Short-term changes over the early evening period in 2008 were found to be significant for both T_{em} and α -pinene emissions (Table 3), the highest values being found at 20:10 (Fig. 3a–e). The emission profile of all 12 trees analyzed in the 2006 daytime sampling remained the same in the 2008 early evening sampling. Furthermore, seven out of nine trees had the same emission profile in the two 2006 samplings.

3.2 Comparing foliar monoterpene emission between *Quercus suber* highly infested and not infested by *Cerambyx welensii*

The limonene chemotype was the most abundant in trees highly infested by *C. welensii* overall (85.7 to 88.9% of trees, depending on the sampling time) (Table 1). None of the trees not infested by *C. welensii* showed the limonene chemotype in the daytime sampling, and this profile was less common in these trees than in trees highly infested by *C. welensii* in the early evening (58.8%) and late evening (66.6%) samplings (G-test: $G_1 = 4.06$, $P = 0.04$; $G_1 = 0.34$, $P = 0.56$, respectively). Differences in individual monoterpene relative emissions between the two study groups were significant only in the 2006 daytime sampling (Table 1, Fig. 1). Specifically, infested trees had both lower percentages of the three pinene-type compounds (1.85 times on average; LMM: $\chi^2_1 \geq 4.89$, $P \leq 0.03$, for all three compounds) and higher percentages of limonene (4.28 times; LMM: $\chi^2_1 = 6.27$, $P = 0.01$) than non-infested trees. Differences in emission composition (relative emissions for limonene, myrcene, and pinene-type compounds together) between the two study groups were significant in the 2006 daytime and 2008 early evening samplings (PERMANOVA: $F_1 = 5.14$, $P = 0.04$ and $F_1 = 4.45$, $P = 0.03$, respectively) but not in 2008 late evening sampling ($F_1 = 0.02$, $P = 0.94$).

Fig. 2 Representative chromatogram of foliage monoterpenes from two *Quercus suber* trees, one highly infested by *Cerambyx welensii*, and another one not infested by this woodborer. Compounds identified: (1) α -thujene; (2) α -pinene; (3) camphene; (4) sabinene; (5) β -pinene; (6) myrcene; (7) α -phellandrene; (8) limonene; (9) cineol; (10) ocimene; (11) γ -terpinene

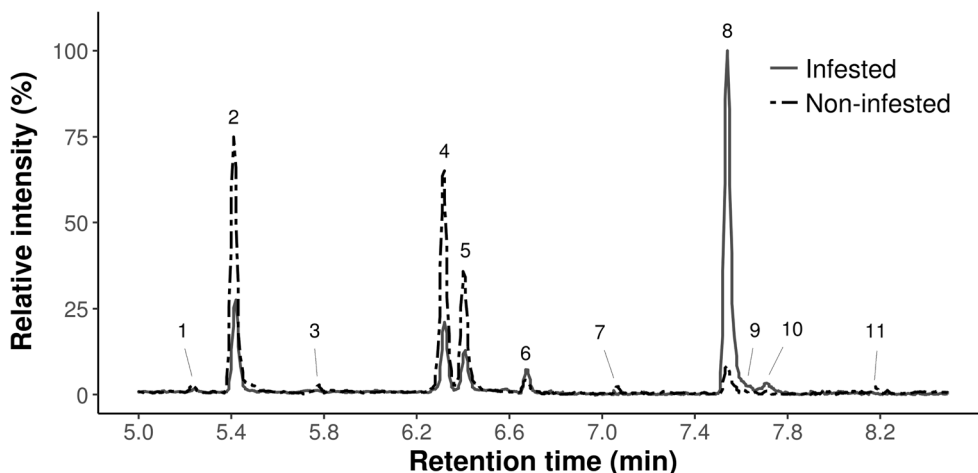


Table 3 Linear mixed model results (*P* values only) of individual monoterpene and total monoterpene emissions (T_{em} , the emission of the five main monoterpenes pooled) of *Quercus suber* trees over the earlyevening (19:00–21:20 h) in 2008, for main effects of measurement time and infestation by *Cerambyx welensii*

Source	<i>P</i> values					
	α -Pinene	Sabinene	β -Pinene	Myrcene	Limonene	Total (T_{em})
Measurement time	<i>0.02</i>	<i>0.09</i>	<i>0.07</i>	0.14	0.25	<i>0.03</i>
Infestation	0.93	0.83	0.77	0.93	<i>0.05</i>	0.21
Measurement time \times infestation	<i>0.03</i>	<i>0.03</i>	<i>0.01</i>	0.57	0.20	0.53

P values < 0.1 are shown in italics

The overall T_{em} values obtained from all three sampling times showed no significant differences between the two study groups (LMM: $\chi^2_1 \leq 1.56$, $P \geq 0.15$) (Table 1 and Fig. 3f). Limonene emissions from infested trees significantly exceeded that of non-infested trees in the 2006 daytime (943.4 ng gdw⁻¹ h⁻¹ and 170.6 ng gdw⁻¹ h⁻¹, respectively; LMM: $\chi^2_1 = 14.23$, $P < 0.001$) and the 2008 early evening sampling (546.3 ng gdw⁻¹ h⁻¹ and 237.4 ng gdw⁻¹ h⁻¹, respectively; LMM: $\chi^2_1 = 3.59$, $P = 0.05$) (Fig. 3f). Further, there was a significant measurement time \times infestation interaction effect on the emission of the three pinene-type compounds in 2008 (Table 3), the largest difference in summed emissions of the three compounds (32.2 times higher in non-infested trees) being observed at 19:35 h (Fig. 3b).

4 Discussion

Plants invest a great deal in volatile terpene production and emission, and multiple stresses affect both constitutive and induced volatile emissions, influencing plant-insect relationships (Loreto et al. 2014; Maja et al. 2014). The ecological function of both volatile production and intraspecific diversification (presence of chemotypes) is not well understood, though the latest research suggests an indirect role in improving thermotolerance and protection against oxidants (Niinemets et al. 2013; Loreto et al. 2014). Under summer conditions, we found differences in foliar monoterpene emission between *Q. suber* trees highly infested by the woodborer *C. welensii* and neighboring non-infested trees, in terms of both emission rate and composition. The differences were found both in the daytime and the early evening periods, with a high presence of limonene in emissions by infested trees and pinene-type compounds dominating emissions by non-infested trees.

The main foliar monoterpenes we found from *Q. suber* were limonene, α -pinene, β -pinene, sabinene, and myrcene (accounting on average for 87.2 to nearly 100% of the total emissions). The same five compounds have previously been found in other regions to represent up to 90% of total

emissions from *Q. suber* (Staudt et al. 2004; Pio et al. 2005). Poor soil water availability, high temperatures, and low relative humidity, conditions that are common during summer in the south of the Iberian Peninsula, may determine physiological responses of vegetation that affect monoterpene emission (Niinemets et al. 2002; Lavoit et al. 2011). In Spain, environmental conditions similar to those we found during our sampling periods (T , 33.5–40.5 °C; RH , < 40%) caused a drop in *Q. ilex* emissions (Núñez et al. 2002). Under summer conditions, we observed low daytime emission rates (on average, 1.84 $\mu\text{g gdw}^{-1} \text{h}^{-1}$) that were within the range of 1 to 5 $\mu\text{g gdw}^{-1} \text{h}^{-1}$ previously reported from *Q. suber* (Pio et al. 2005) and *Q. ilex* (Núñez et al. 2002). A high occurrence of the genotype-dependent limonene chemotype (up to 62% of trees) has been reported in *Q. suber* (Staudt et al. 2004). In our results, the limonene profile was found in < 67% in *Q. suber* trees not infested by *C. welensii* but was much more common in trees infested by *C. welensii* (> 85% of trees, regardless of the year and sampling time).

Stress-associated changes in emission may occur even on a relatively short time scale of days to hours (Pio et al. 2005; Loreto et al. 2014). Moreover, during the early evening and late evening periods, selected for their relevance to *C. welensii* behavioral patterns, emission rates were probably affected by low *PPFD* values. The low repeatability estimates we found for measured emission rates could suggest that they were affected by short-term environmental and physiological constraints (e.g., light, temperature, photosynthesis); in contrast, the high within-individual repeatability in emission composition we found could be attributable to systematic factors (e.g., genotype, constant environmental conditions, and physiological status of trees). Across the early evening period, differences we found in the effect of measurement time on the emissions of pinene-type compounds and limonene could suggest a higher dependence of the former on variables such as temperature, light, or water stress (Núñez et al. 2002; Grote 2007). Physicochemical characteristics of monoterpenes (solubility, volatility, and diffusivity) can control emissions and interact with physiological limitations (Niinemets et al. 2013). The increase in the relative contribution of limonene to

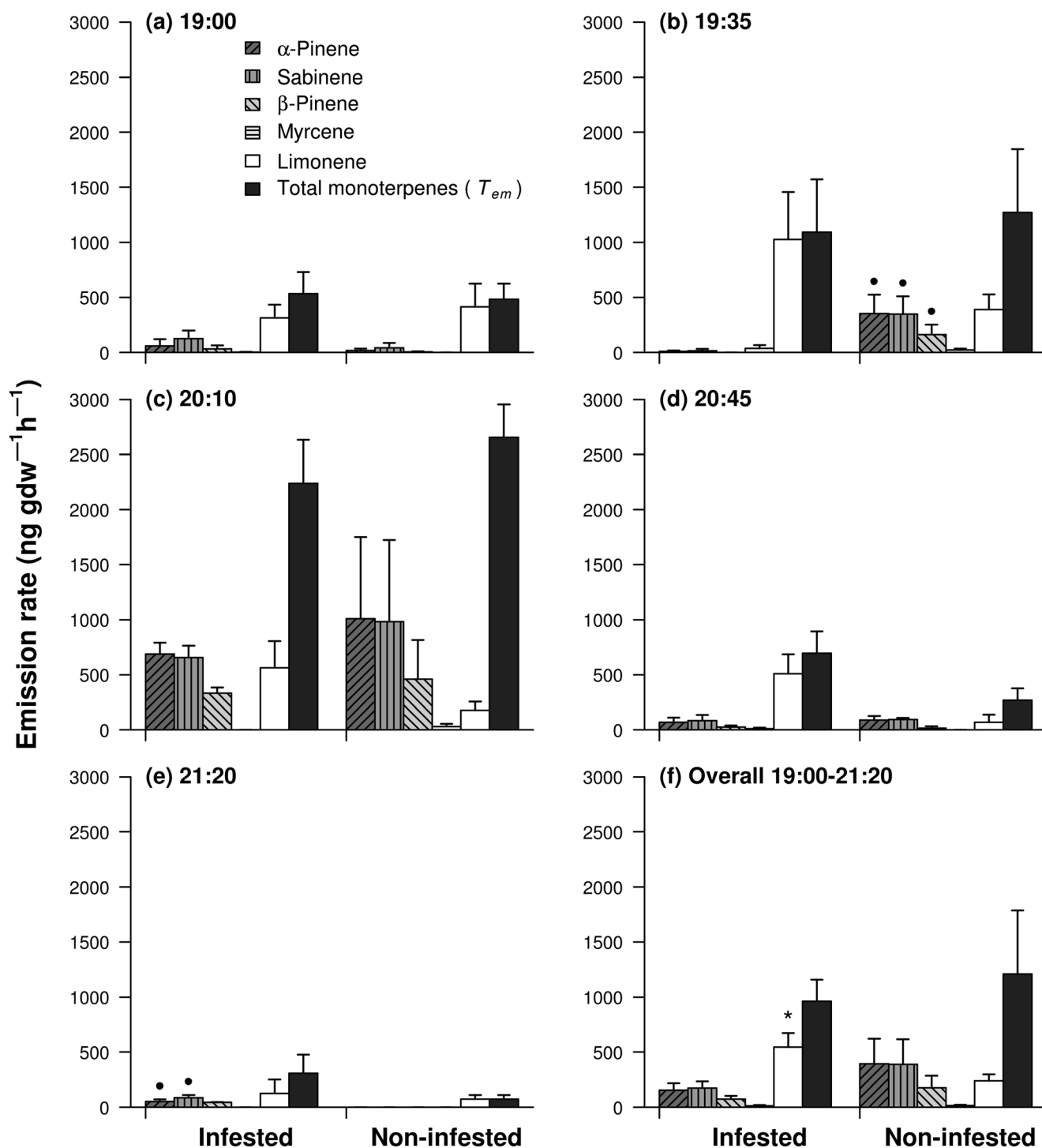


Fig. 3 Short-term variation, during 2008 early evening, in the mean (\pm SE) emission rates ($\text{ng gdw}^{-1} \text{h}^{-1}$) of individual monoterpenes and total monoterpene emission (T_{em} , the emission of the five compounds pooled) from *Quercus suber* stratified by whether or not trees were infested by

Cerambyx welensii. For each measurement time (panels a–f), a symbol over a bar denotes significant differences between trees with different infestation status (*, $P < 0.05$; •, $P < 0.1$. LMM with Benjamini-Hochberg correction. $n = 3\text{--}4$ per measurement time for each infestation status)

Q. suber emissions during the early hours of the night has been explained by storage of a certain amount of it inside leaves, together with its low volatility compared with pinene-type compounds (Pio et al. 2005). It has been suggested that non-stored constitutive volatiles play a role in the

mitigation of some abiotic stresses, by stabilizing membranes and serving as antioxidants (see, for example, Niinemets et al. 2013). Nevertheless, the role of monoterpenes, especially limonene, and other isoprenoids in such defense mechanisms remains unclear (Loreto et al. 2014).

Volatile chemicals released by damaged or stressed host plants may provide cues that cerambycids use to locate larval hosts (Allison et al. 2004; Millar and Hanks 2017), and visual cues could also influence host choice (Torres-Vila et al. 2017); however, the underlying host-selection mechanisms by long-horned wood-boring beetles remain poorly understood. In our study area, Lopez-Pantoja et al. (2008) reported a continuous exchange of *C. welensii* individuals with neighboring areas, suggesting a host-selection mechanism that could mediate these movements. *C. welensii* has been found to show antennal sensitivity to limonene, myrcene, and pinene-type compounds (Sánchez-Osorio 2005). Nevertheless, monoterpenes (in particular, β -pinene) have been found to produce a low attraction of *C. welensii* to traps, compared with high-release-rate synthetic blends mimicking fermenting plant material (Sánchez-Osorio et al. 2016).

Our experimental design did not allow us to establish a causal relationship between *C. welensii* activity and monoterpene emission, as the influence of *C. welensii* colonization on emissions was not the primary focus of this study. Moreover, other attractive compounds (ethanol and other compounds from decaying tissues and bark exudates) could act either alone or in combination with foliar emissions. Nonetheless, the differences we found in the presence of the two main emission profiles (limonene-type and pinene-type) between infested and non-infested trees, especially during the early evening period, point to the influence of the emission profile on host choice by an insect with crepuscular flight activity. Moreover, the low emission rates we found in the late evening period suggest that the key time for discriminating between trees via monoterpene detection is the early evening period, near dusk, when *C. welensii* adults become active.

5 Conclusion

This study has evidenced differences in foliar monoterpene emission between *Q. suber* trees highly infested by the woodborer *C. welensii* and neighboring cork oak trees free of this cerambycid, in both daytime and early evening periods. Limonene dominated in trees infested by *C. welensii*, whereas pinene-type compounds did in non-infested trees. Given these results, we hypothesize that the detection—especially during the time when *C. welensii* adults initiate daily flight—of certain compounds (e.g., limonene), together with the detection of specific ratios of several monoterpenes (e.g., those of limonene to pinene-type compounds), has a role in the intraspecific host selection by *C. welensii*. These results contribute to understand plant-insect interactions, specifically those affecting host selection by *C. welensii* in *Q. suber* stands and may have important implications in the integrated and sustained pest management of this species. Further research should be warranted to assess the way changes in physiological

performance of *Q. suber* (due to both woodborer damage and environmental stress) influence the emission behavior and the suitability of trees for *C. welensii* larval development, as well as to test the effect of some monoterpenes (mainly, limonene and pinene-type compounds, especially when they are combined with either moderate- or high-release-rate synthetic blends) on the attraction of *C. welensii* in laboratory and the field studies. Additional studies, using artificially baited trees and/or artificially weakened trees without larval activity, could be also warranted to assess whether emissions affect host selection.

Acknowledgments We thank Sebastiana Malia, Agustín Rincón, and María del Mar González for their assistance in both the field and the laboratory, and Dr. Manuel Fernández for his constructive comments.

Statement on data availability The datasets generated and/or analyzed during the current study are available in the Open Science Framework (Sánchez-Osorio et al. 2019) at <https://doi.org/10.17605/OSF.IO/D2ZH3>.

Compliance with ethical standards

Conflict of interest The authors declare that they have no conflict of interest.

References

- Alejano R, Domingo JM, Fernández M (coords) (2011) Manual para la gestión sostenible de las dehesas andaluzas. Foro para la Defensa y Conservación de la Dehesa "Encinal". Universidad de Huelva
- Allison JD, Borden JH, Seybold JH (2004) A review of the chemical ecology of the Cerambycidae. *Coleoptera* 14:123–150
- Aronson J, Pereira JS, Pausas JG (eds) (2009) Cork oak woodlands on the edge. Ecology, adaptive management, and restoration. Society for Ecological Restoration International, Island Press, Washington
- Bates D, Maechler M, Bolker B, Walker S (2014) lme4: linear mixed-effects models using Eigen and S4. R package version 1.1–7. <http://CRAN.R-project.org/package=lme4>
- Falconer DS, Mackay TFC (1996) Introduction to quantitative genetics, fourth ed. Longman Group Limited, Harlow
- Grote R (2007) Sensitivity of volatile monoterpene emission to changes in canopy structure: a model-based exercise with a process-based emission model. *New Phytol* 173:550–561
- ICNF (2013) IFN6 – Áreas dos usos do solo e das espécies florestais de Portugal continental. Resultados preliminares Instituto da Conservação da Natureza e das Florestas, Lisboa, 34 pp
- Lavoir AV, Duffet C, Mouillot F, Rambal S, Ratte JP, Schnitzler JP, Staudt M (2011) Scaling-up leaf monoterpene emissions from a water limited *Quercus ilex* woodland. *Atmos Environ* 45:2888–2897
- Lenth RV (2014) lsmmeans: least-squares means. **R package version 2.10**. <http://CRAN.R-project.org/package=lsmmeans>
- López-Pantoja G, DomínguezNevado L, Sánchez-Osorio I (2008) Mark-recapture estimates of the survival and recapture rates of *Cerambyx welensii* Küster (Coleoptera Cerambycidae) in a cork oak dehesa in Huelva (Spain). *Cent Eur J Biol* 3:431–441
- Loreto F, Pollastri S, Fineschi S, Velikovic V (2014) Volatile isoprenoids and their importance for protection against environmental constraints in the Mediterranean area. *Environ Exp Bot* 103:99–106
- Maja MM, Kasurinen A, Yli-Pirilä P, Joutsensaari J, Klemola T, Holopainen T, Holopainen JK (2014) Contrasting responses of

- silver birch VOC emissions to short- and long-term herbivory. *Tree Physiol* 34:241–252. <https://doi.org/10.1093/treephys/tpt127>
- Martín J, Cabezas J, Buyolo T, Patón D (2005) The relationship between *Cerambyx* spp. damage and subsequent *Biscogniauxia mediterranea* infection on *Quercus suber* forests. *Ecol Manag* 216:166–174
- Millar JG, Hanks LM (2017) Chemical ecology of cerambycid beetles. In: Wang Q (ed) *Cerambycidae of the world: biology and management*. CRC Press/Taylor & Francis, Boca Raton
- Niinemets U, Seufert G, Steinbrecher R, Tenhunen JD (2002) A model coupling foliar monoterpene emissions to leaf photosynthetic characteristics in Mediterranean evergreen *Quercus* species. *New Phytol* 153:257–275
- Niinemets U, Kännaste A, Copolovici L (2013) Quantitative patterns between plant volatile emissions induced by biotic stresses and the degree of damage. *Front Plant Sci* 4:262. <https://doi.org/10.3389/fpls.2013.00262>
- Núñez L, Plaza J, Pérez-Pastor R, Pujadas M, Gimeno B, Bermejo V, García-Alonso S (2002) High water vapour pressure deficit influence on *Quercus ilex* and *Pinus pinea* field monoterpene emission in the central Iberian Peninsula (Spain). *Atmos Environ* 36:4441–4452
- Oksanen F, Blanchet G, Kindt R et al. (2015) *Vegan: community ecology package*. R package version 2.2–1. <http://CRAN.R-project.org/package=vegan>
- Pio CA, Silva PA, Cerqueira MA, Nunes TV (2005) Diurnal and seasonal emissions of volatile organic compounds from cork oak (*Quercus suber*) trees. *Atmos Environ* 39:1817–1827
- Sallé A, Nageleisen LM, Lieutier F (2014) Bark and wood boring insects involved in oak declines in Europe: current knowledge and future prospects in a context of climate change. *Ecol Manag* 328:79–93. <https://doi.org/10.1016/j.foreco.2014.05.027>
- Sánchez-Osorio I (2005) Orientación olfativa de *Cerambyx welensii* Küster y *Prinobius germari* Dejean, principales cerambicidos xilófagos de encina (*Quercus ilex* L. subsp. *ballota*) y alcornoque (*Quercus suber* L.), para la localización de hospedantes. Doctoral Thesis, University of Huelva, Huelva
- Sánchez-Osorio I, López-Pantoja L, Paramio AM, Lencina JL, Gallego D, Domínguez L (2016) Field attraction of *Cerambyx welensii* to fermentation odors and host monoterpenes. *J Pest Sci* 89:59–68
- Sánchez-Osorio I, López-Pantoja G, Tapias R, Pareja-Sánchez E, Domínguez L (2019). Monoterpene emission of *Quercus suber* L. highly infested by *Cerambyx welensii* Küster. V 20 Aug 2019. Open Science Framework [Dataset] <https://doi.org/10.17605/OSF.IO/D2ZH3>
- Staudt M, Mir C, Joffre R, Rambal S, Bonin A, Landais D, Lumaret R (2004) Isoprenoid emission of *Quercus* spp. (*Q. suber* and *Q. ilex*) in mixed stands contrasting in interspecific genetic introgression. *New Phytol* 163:573–584
- Tiberi R, Branco M, Bracalini M, Croci F, Panzavolta T (2016) Cork oak pests: a review of insect damage and management. *Ann For Sci* 73: 219–232. <https://doi.org/10.1007/s13595-015-0534-1>
- Torres-Vila LM, Sanchez-González A, Ponce-Escudero F, Martín-Vertedor D, Ferrero-García JJ (2012) Assessing mass trapping efficiency and population density of *Cerambyx welensii* Küster by mark-recapture in dehesa open woodlands. *Eur J Forest Res* 131: 1103–1116. <https://doi.org/10.1007/s10342-011-0579-0>
- Torres-Vila LM, Mendiola-Díaz FJ, Conejo-Rodríguez Y, Sánchez-González Á (2016) Reproductive traits and number of matings in males and females of *Cerambyx welensii* (Coleoptera: Cerambycidae) an emergent pest of oaks. *Bull Entomol Res* 106: 292–303
- Torres-Vila LM, Mendiola-Díaz FJ, Sánchez-González Á (2017) Dispersal differences of a pest and a protected *Cerambyx* species (Coleoptera: Cerambycidae) in oak open woodlands: a mark-recapture comparative study. *Ecol Entomol* 42:18–32. <https://doi.org/10.1111/een.12355>
- Vives E (2000) Coleoptera, Cerambycidae. *Fauna Ibérica*, vol 12 (ed. by M. A Ramos et al.), Museo Nacional de Ciencias Naturales. CSIC, Madrid
- Zhang Q-H, Birgersson G, Zhu J, Lofstedt C, Lofqvist J, Schlyter F (1999) Leaf volatiles from nonhost deciduous trees: variation by tree species, season and temperature, and electrophysiological activity in *Ips typographus*. *J Chem Ecol* 8:1923–1943

Publisher's note Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.